

1 **Modified local sands for the mitigation of harmful algal blooms**

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10 ABSTRACT

11 A new method was developed for marine harmful algal bloom (HAB) mitigation
12 using local beach sand or silica sand modified with chitosan and polyaluminum
13 chloride (PAC). Untreated sand was ineffective in flocculating algal cells, but 80%
14 removal efficiency was achieved for *Amphidinium carterae* Hulburt and a *Chlorella*
15 sp. in 3 min ($t_{80} = 3$ min) using 120 mg L⁻¹ sand modified with 10 mg L⁻¹ PAC and 10
16 mg L⁻¹ chitosan. After several hours 92% – 96% removal was achieved. The t_{80} for
17 removing *A. carterae* using the modifiers only (PAC and chitosan combined) was 60
18 min and for *Chlorella* sp. 120 min, times which are much slower than with the
19 corresponding modified sand. Sands were critical for speeding up the kinetic
20 processes of flocculation and sedimentation of algal flocs. PAC was helpful in
21 forming small flocs and chitosan is essential to bridge the small flocs into large dense
22 flocs. Chitosan was also important in inhibiting the escape of cells from the flocs.
23 Chitosan and PAC used together as modifiers make it possible to use local beach
24 sands for HAB mitigation in seawater. Economical and environmental concerns could
25 be reduced through the use of sands and biodegradable chitosan, but the potential
26 impacts of PAC need further study.

27 *Keywords:* Harmful algal bloom; Seawater; Modified sands; Chitosan; Polyaluminum
28 chloride (PAC); Synergistic effect.

29

30 1. Introduction

31 Harmful algal blooms (HABs) pose a serious threat to public health, aquatic

32 organisms, commercial fisheries, and the quality of freshwater lakes, rivers and
33 reservoirs, as well as marine coastal environments. Over the past decade, there has
34 been increasing interest in bloom mitigation strategies, though progress towards field
35 applications has still been slow (Anderson, 1997). Significant attention has been
36 focused on the use of clays as a means to remove HAB cells from the water column
37 through flocculation and sedimentation. Many of these experiments were laboratory
38 based (Beaulieu et al., 2005; Pan et al., 2006a; Pierce et al., 2004; Sengco et al., 2001;
39 Yu et al., 1994), with some field demonstrations in Japan (Shirona, 1989),
40 Australia (Atkins et al., 2001), China (Pan et al., 2006b) and South Korea (e.g., Lee et
41 al., 2008). The environmental impacts of clay flocculation are generally positive,
42 though there are studies that document negative effects. On the positive side, clay
43 flocculation had little or no effect on marine organisms such as juvenile clams, fish,
44 and invertebrates (Lewis et al. 2003; Archambault et al, 2004; Sengco and Anderson,
45 2004). In one of these studies, however, a growth effect on juvenile hard clams was
46 observed (compared to no-clay controls) with clay maintained in suspension for two
47 weeks. These results suggest that clay applications in the field are likely more
48 detrimental to clams under flow conditions leading to prolonged in situ resuspension
49 of clay than under conditions that promote rapid sedimentation. Shumway et al. (2003)
50 also report negative impacts on filter-feeding invertebrates using relatively high levels
51 of clay. The magnitude of impacts is thus dependent on the flow regime, duration of
52 exposure to resuspended clay, and the total clay loading.

54 However, clays are not immediately available at some locations that have HAB
55 problems, and transportation costs may render this method uneconomical. There is
56 also a common ecological concern about the dumping of large amounts of exotic
57 materials into aquatic systems. As an alternative strategy, the use of native ecological
58 materials such as local beach sands or soil (that naturally enter the aquatic system
59 through rivers or rainfall) could in principle minimize the costs and ecological risk to
60 aquatic environments. Sands, however, have markedly different physical
61 characteristics from clays, and by themselves, will not flocculate and remove HAB
62 cells.

63 In freshwater HAB mitigation, Pan and co-workers found that local soil particles
64 including sands can be highly effective in removing cyanobacterial cells and
65 improving water quality, but only after modification using small amounts of a natural,
66 biodegradable material called chitosan (Pan et al., 2006b; Zou et al., 2006; Pan et al.,
67 2011). These authors found that the polymeric netting and bridging function of
68 chitosan was the key mechanism that allowed local soil particles to be highly effective
69 in flocculating HAB cells. In this approach, the chitosan made a "net" that captured
70 the HAB cells and other particles, and the soils provided the ballast or mass to carry
71 the aggregates to the bottom. These encouraging results in freshwater have, however,
72 limited direct applicability in marine systems, as high ionic strength and alkalinity
73 prevent the unfolding of the polymer chain, thereby weakening chitosan's netting and
74 bridging properties (Qun and Ajun, 2006; Zou et al., 2005).

75 Polyaluminum chloride (PAC), a commonly used inorganic coagulant, is highly

76 effective in potable water treatment where it is used routinely to flocculate and
77 remove suspended particles. PAC has been tested in marine systems and has been
78 shown to reduce the amount of clays needed to remove HAB organisms (Pierce et al.,
79 2004; Sengco et al., 2001; Yu et al., 1994). The addition of PAC increases the
80 chemical affinity of clay surfaces. According to laboratory studies, however, algal cell
81 flocculation by clays plus PAC was temporary (Sengco et al., 2001; Sun et al., 2004).
82 Most of the cells could escape from the flocs and resume their growth. Motile
83 dinoflagellate species were thus more difficult to be removed permanently through
84 flocculation compared to non-motile diatoms (Yu et al., 1994), indicating that motility
85 was an important factor affecting bloom mitigation through clay flocculation.
86 Furthermore, the PAC floc was light, which did not settle easily or was resuspended
87 with only modest currents (Beaulieu et al. 2005).

88 No efforts have been made thus far to use local beach sands to irreversibly
89 flocculate and sediment marine HAB cells. Here, a modification of the approach to
90 suppress freshwater HABs using local beach sands and polymers was developed for
91 algal bloom mitigation in seawater. The synergistic effects of chitosan and PAC
92 (hereafter termed "modifiers") with two types of sands were investigated for the
93 removal of *Amphidinium carterae* and *Chlorella* sp. The results demonstrate that it is
94 possible to use modified local or commercially available sands to irreversibly remove
95 a high percentage of the two types of HAB cells from seawater.

96 2. Materials and Methods

97 2.1. Algal species and culture

98 Two algal species were used - *Amphidinium carterae* Hulburt, a motile
99 dinoflagellate, and a marine *Chlorella* sp. which is very small, and non-motile. *A.*
100 *carterae* is considered a HAB species because of its production of haemolysins, and it
101 has also been linked to fish mortalities(Hulburt, 1957; Yasumoto et al., 1987).
102 Although *Chlorella* is not listed as a harmful species on some lists, it is known for its
103 ability to produce dense blooms that can have adverse consequences, such as the
104 decimation of the oyster industry on Long Island following eutrophication stimulated
105 by duck farm effluents (Ryther, 1954). *A. carterae* was obtained from Oceanography
106 College, Ocean University of China and *Chlorella* sp. was supplied by Seaweed
107 Inheritance Breeding Center of Shandong Oriental Ocean Sci.-Tech. Co. Ltd..

108 The cells were grown in *f/2* medium (Guillard and Hargraves, 1993) made with
109 synthetic seawater. The synthetic seawater was composed of 23.939 g L⁻¹ NaCl, 5.079
110 g L⁻¹ MgCl₂·6H₂O, 3.994 g L⁻¹ Na₂SO₄, 1.123 g L⁻¹ CaCl₂, 0.667 g L⁻¹ KCl, 0.196 g
111 L⁻¹ NaHCO₃, 0.098 g L⁻¹ KBr, 0.027 g L⁻¹ H₃BO₃, 0.003 g L⁻¹ NaF and 0.024 g L⁻¹
112 SrCl₂·6H₂O. The medium was adjusted to pH 8.2 before autoclaving by adding either
113 0.1 mol L⁻¹ NaOH or 0.1 mol L⁻¹ HCl solutions. Algal batch cultures were maintained
114 at 25±1°C under continuous cool white fluorescent light of 2000-3000 lux on a 12h
115 light and 12h darkness regimen in the illuminating incubator (LRH-250-G,
116 Guangdong Medical Apparatus Co. Ltd., China).

117 2.2. Sands and modifiers

118 Two kinds of sand were used. One was SiO₂ (silica sand) analytical grade,
119 purchased from Sinopharm Chemical reagent Co., Ltd.. Another was local sand which
120 collected from a Yellow Sea beach in Yantai, China. The two sands were washed with
121 deionized water, dried at 100°C, and sieved through 180 mesh (<90 μm).

122 Chitosan was obtained from Qingdao Haisheng Bioengineering Co. Ltd. The
123 chitosan flakes were dissolved by adding 100 mg chitosan to 10 mL of 0.5% HAc and
124 stirring until all the chitosan was dissolved. This solution was diluted with deionized
125 water to obtain a final concentration of 1mg mL⁻¹ before use (Zou et al., 2006). PAC
126 was supplied by Dagang Reagent Plant, Tianjin, China. The basicity (B= [OH]/ [Al])
127 of PAC was 2.4 and its Al₂O₃ content was 30%. The PAC was dissolved in deionized
128 water to obtain a solution of 1 mg mL⁻¹. The chitosan and PAC solutions were
129 prepared freshly before each set of experiments.

130 2.3. Algal flocculation

131 Flocculation experiments were conducted using a jar test apparatus (ZR3-6,
132 Zhongrun Water Industry Technology Development Co. Ltd., China) using cultures in
133 mid- to late-exponential growth phase. The initial cell concentrations of *A. carterae*
134 and *Chlorella* sp. were 3.25 - 3.42×10⁵ cells mL⁻¹ and 6.65 - 6.82×10⁶ cells mL⁻¹,
135 respectively. Two hundred milliliters of experimental culture were transferred into a
136 250 mL beaker, stirred at 200 rpm for 2 min, followed by 30 rpm for another 5 min.
137 Chitosan alone, PAC alone, chitosan plus PAC together, and chitosan plus PAC plus
138 sands were added to the algal culture in different flocculation experiments. The

139 control culture was run without adding any sands or modifiers.

140 Samples from 2 cm below the surface of the experimental beaker were collected
141 after sedimentation at different times and the cells enumerated in a counting chamber
142 under an electromotive microscope (Axioskop 2 mot plus, Carl ZEISS, Germany)
143 after being fixed by Lugol solution. The removal efficiency of cells was calculated as
144 $(\text{initial cell concentration} - \text{sample cell concentration}) / \text{initial cell concentration} \times$
145 100%. Algal flocs were collected by pipette and observed under the microscope.

146 Algal floc size and size distribution during the flocculation process were monitored
147 with a laser particle size analyzer Mastersizer 2000 (Malvern Co. United Kingdom).
148 The culture was drawn into the Mastersizer and back to the jar by a peristaltic pump
149 (BT00-300M, Baoding Longer Precision Pump Co. Ltd., China) at a flow rate of 34
150 mL min⁻¹ (Zhang et al., 2007). Samples were at the same position in the jar, which was
151 located between the impeller and the top of suspension. Algal floc size was denoted
152 by the measured mean diameter (d_{50}).

153 2.4. Viability and growth of algae after flocculation

154 The effect of PAC or chitosan with PAC on the viability and the growth of *A.*
155 *carterae* after flocculation was investigated using two strategies. In the first
156 experiment, fresh f/2 medium was added to the supernatant without disturbing the
157 algal flocs (Sengco et al., 2001; Sun and Choi, 2004). This flask was maintained in an
158 illuminated incubator, and viability and growth of the cells were monitored by
159 measuring the cell concentrations in the supernatant after 24 and 48 hours. In the
160 second experiment, flocs were maintained in the incubator without fresh f/2 medium

161 or light.

162 **3. Results**

163 3.1. Algal flocculation using modified sands

164 Compared with control experiments, 100 mg L⁻¹ silica sand or local sand was
165 ineffective in removing *A. carterae* and *Chlorella* sp. (Fig.1). However, sands
166 modified using chitosan and PAC combined were highly efficient in flocculating and
167 sinking algal cells. The removal efficiency with 120 mg L⁻¹ modified sands containing
168 10 mg L⁻¹ chitosan and 10 mg L⁻¹ PAC reached 80% for the two algal species within 3
169 min (t_{80} =3 min), whereas the removal efficiencies of only 10 mg L⁻¹ chitosan plus 10
170 mg L⁻¹ PAC on *A. carterae* (Fig.1A) and *Chlorella* sp. (Fig.1B) were 54% and 43%,
171 respectively. The t_{80} of the modifiers alone for *A. carterae* removal was 60 min and
172 that for *Chlorella* sp. was 120 min. Using only sands, the removal efficiencies of *A.*
173 *carterae* and *Chlorella* sp. after 240 min were 26% and 7% (Figs. 1A, 1B). This
174 increased to 96% and 92% when the chitosan and PAC modifiers were added with the
175 sand. The results in Fig.1 also demonstrate that there was no large difference between
176 silica sand and local beach sand on HAB cell removal if the modifiers chitosan and
177 PAC were present.

178 3.2. Synergistic effect of chitosan and PAC on algal cell removal

179 When chitosan was used alone, cell removal efficiencies increased with increasing
180 dosage of chitosan (0 – 20 mg L⁻¹ for *A. carterae* and 0 – 50 mg L⁻¹ for *Chlorella* sp.;
181 Fig.2). However, the removal efficiency of *A. carterae* (Fig.2A) was maximally 71%
182 at 20 mg L⁻¹ chitosan and that of *Chlorella* sp. (Fig.2B) was only 51% at 50 mg L⁻¹,

183 which suggests that chitosan is not as efficient at removing algal cells from seawater
184 as it is in fresh water (Pan et al., 2006b; Zou et al., 2006).

185 Cell removal efficiency for both species increased when PAC and chitosan were
186 used together (Fig. 2). After the addition of 5 mg L⁻¹ PAC with 10 mg L⁻¹ chitosan, the
187 removal efficiency of *A. carterae* and *Chlorella* sp. increased to 92% and 62% from
188 68% and 11%, respectively. When 10 mg L⁻¹ PAC was added with 10 mg L⁻¹ chitosan,
189 the *A. carterae* removal efficiency increased by an additional 28% over that with
190 chitosan alone, and that of *Chlorella* sp. increased by 78%.

191 3.3. Synergistic effect of chitosan and PAC on algal floc formation

192 The formation and development of algal flocs using 10 mg L⁻¹ PAC or PAC with 10
193 mg L⁻¹ chitosan were investigated using *Chlorella* sp. as the target species. The floc
194 size (Fig. 3A) and size distributions (Fig. 3B) were monitored. Compared with PAC
195 alone, the algal flocs of PAC plus chitosan increased in size much faster in the first
196 two minutes. During the slow stir phase, algal floc size increased to a plateau. The
197 floc size of PAC plus chitosan increased to 860 µm, compared to that of PAC alone,
198 for which the size was approximately 600 µm. The floc produced by chitosan and
199 PAC appeared rapidly and quickly increased in size to form larger particles than with
200 PAC only.

201 At 7 min, the stir was over and floc size distribution curves were shown in Fig. 3B.
202 The floc size distribution of PAC alone ranged between 316 µm and 1259 µm, with
203 the highest peak at 631 µm. The size distribution of PAC plus chitosan was between
204 417 µm and 2188 µm, with the highest peak at 955 µm.

205 3.4. Synergistic effect of chitosan and PAC on cell viability

206 An experiment examining the synergistic effect of chitosan and PAC on the viability
207 and growth of *A. carterae* was divided into three treatments: (1) 10 mg L⁻¹ PAC only,
208 (2) 10 mg L⁻¹ PAC plus 10 mg L⁻¹ chitosan, (3) 10 mg L⁻¹ PAC plus 20 mg L⁻¹
209 chitosan. After these flocculation experiments, the residual cell concentration in the
210 supernatant of the three treatments was 1.2 - 1.6×10⁴ cells mL⁻¹, approximately 4% of
211 the original concentration prior to the treatment. The cell concentration for all the
212 treatments roughly doubled to 2.8 - 3.0×10⁴ cells mL⁻¹ after 24 hours of incubation in
213 an incubator with light and added nutrients (Fig. 4A). After another 24 hours, the cell
214 concentration with PAC only increased dramatically to 12.4 ×10⁴ cells mL⁻¹, while the
215 concentration in the treatments of PAC plus 20 mg L⁻¹ chitosan rose to 5.05 ×10⁴ cells
216 mL⁻¹, approximately half of the concentration with PAC only.

217 The results shown in Fig.4B demonstrate that the cell concentration in the
218 supernatant of the three treatments in the incubator with no light or added nutrients
219 decreased gradually throughout the study interval. However, the algal cell
220 concentrations of PAC plus chitosan used together were less than that of PAC alone
221 and the cell concentration was inversely related to the chitosan dosage. After 28 days,
222 the concentration of algal cells in supernatant was only 300 cells mL⁻¹, indicative of
223 almost no recovery of *A. carterae* cells under conditions similar to those found near
224 bottom sediments.

225 **4. Discussion**

226 In this study, a method was developed that uses sands or local soils that could be

227 collected from the immediate vicinity of a HAB, and used in conjunction with small
228 amount of chitosan and PAC to flocculate and effectively remove cells from the water
229 column. Our results demonstrate that PAC was needed to maintain the netting and
230 bridging function of chitosan in seawater and to form small flocs, while chitosan was
231 essential in bridging the small flocs into large and dense flocs that hindered the escape
232 of cells from the flocs. As the safe and cheap carrier of these modifiers, sand was
233 critical for speeding up sedimentation. This approach, which was a modification of
234 the one used successfully for HAB removal in freshwater systems (Pan et al., 2006b;
235 Pan et al., 2011), greatly minimizes environmental concerns for mitigation of HABs
236 in seawater using clays since the use of native beach sands has few environmental
237 concerns. As discussed below, however, there are still some issues that need to be
238 addressed if this method is used for field applications on natural blooms.

239 4.1. Synergistic effects of chitosan plus PAC

240 The flocculation of algal cells in natural waters occurs as a result of attractive
241 anion-cation interactions, as well as hydrophobic or polymer interactions (Divakaran
242 and Pillai, 2001; Strand et al., 2002). Sands alone are much less efficient in
243 flocculating algal cells compared to clays such as kaolinite, montmorillonite, and
244 sepiolite (Pan et al., 2006a; Pan et al., 2006b; Pierce et al., 2004; Sengco et al., 2001;
245 Yu et al., 1994). Chitosan and PAC as modifiers increase the surface charge of sands
246 and enhance the netting and bridging interactions with algal cells. Sands also provide
247 the mass or ballast to carry flocs to bottom sediments.

248 Chitosan, a cellulose-like polyelectrolyte biopolymer, is derived from the alkaline

249 deacetylation of crustacean chitin, which possesses several intrinsic characteristics of
250 coagulants and flocculants, i.e., high cationic charge density, long polymer chains,
251 bridging of aggregates and precipitation (Renault et al., 2009; Rinaudo, 2006).
252 Chitosan, by itself, does not flocculate effectively in seawater (Fig. 2). This is because
253 its molecular structure includes abundant amino groups (-NH_2) and hydroxyl groups
254 (-OH) on the chain. The active amine group (-NH_2) of chitosan is easily protonated as
255 -NH_3^+ in dilute acidic solutions, and there is a strong electrostatic repulsion force
256 within and between molecules (Rinaudo, 2006). The high content of positively
257 charged amine groups in the chitosan structure facilitates electrostatic interactions
258 between polymer chains and negatively charged contaminants (Huang et al., 2000;
259 Renault et al., 2009). However, in high ionic strength solutions such as seawater,
260 counter-ions accumulate near the -NH_3^+ group, which would screen the protonated
261 amine groups and decrease the electrostatic repulsion among them (Qun and Ajun,
262 2006; Schatz et al., 2003). This prevents the unfolding of the molecular chain, thereby
263 weakening its netting and bridging properties (Zou et al., 2005).

264 In contrast to chitosan, the high ionic strength of seawater is beneficial to PAC
265 flocculation due to the reduction of the thickness of the electrical double layer which
266 enhances the collision probability of granules. PAC supplies cationic hydrolysis
267 products that are strongly adsorbed on negative particles and can give effective
268 destabilization, leading to the formation of micro-flocs (Renault et al., 2009). Particles
269 with thinner electrical double layers are easier to coagulate because of reduced
270 repulsion. With the high salinity of seawater, flocculation of particles is increased

271 because the thickness of the electrical double layer is decreased due to the
272 compression of the electrolytes (Han and Kim, 2001; Pan et al., 2006b). This explains
273 why PAC is effective in flocculating HAB cells in seawater and why the algal cell
274 removal efficiencies of chitosan are increased remarkably with the addition of PAC.
275 PAC cannot be used by itself in seawater, however, since, discussed by Beaulieu et al.
276 (2005), PAC flocs are light and fluffy and do not settle even in light flow regimes. If
277 these small flocs can be combined and form a stronger, larger, and heavier flocs, then
278 the limitations of PAC flocs can be overcome.

279 The amino groups (-NH₂) and hydroxyl groups (-OH) in chitosan's molecular
280 structure contain single-pair electrons that can offer the electron pair to empty
281 trajectories of metal ions; they then chelate into a complex compound (Bassi et al.,
282 2000). It was reported that there was a positive correlation between chitosan and PAC
283 and the effect of chitosan adsorbing Al³⁺ in solution was very obvious (Zeng et al.,
284 2008). The cationic hydrolysis products of PAC that are adsorbed on the molecule
285 chain of chitosan might increase electrostatic repulsion between them and protonated
286 groups (-NH₃⁺), which would in turn be beneficial to the unfolding of chitosan's
287 molecular chain and weaken the negative effect of high ionic strength on chitosan's
288 netting and bridging properties in seawater. Therefore, PAC and chitosan are
289 complementary in flocculating HAB cells in seawater. Larger and denser algal flocs
290 are formed by the compression of electrical double layer, charge neutralization,
291 adsorption, and netting interactions to bind and bridge cells tightly.

292 4.2. Cell escape from flocs

293 As shown in Figure 4, with light and nutrients provided to cells flocculated using PAC
294 and chitosan alone, cell concentrations in the supernatant doubled in 24 hours, and
295 then doubled again 24 hours later. *Amphidinium* can grow rapidly, with growth rates
296 as high as 2.7 divisions per day (Ismael et al., 1999), so the cell increase in the
297 supernatant of the chitosan plus PAC treatment could be explained entirely by growth
298 with little or no contribution from cells escaping from the flocs. The much larger
299 increase in cell abundance in the PAC only treatment suggests that a significant
300 number of cells escaped into the supernatant.

301 Chitosan flocs were fibrous and formed large entangled masses resembling
302 cobwebs by bridging mechanisms (Fig.5A). The protonated amine group of chitosan
303 attract negatively charged algal cells to produce large and complex flocs that help to
304 prevent the escape of motile cells. In contrast, the flocs of PAC alone were small and
305 there were large numbers of cells around the flocs (Fig. 5B). This implies that PAC
306 does not bridge the algal cells firmly nor bind them as strongly as chitosan does.
307 Overall, the number of cells escaping from the PAC plus chitosan flocs was small, and
308 the method appeared promising for bloom mitigation. The addition of sand would
309 make cell escape even more difficult.

310 4.3 Environmental impacts

311 One of the challenging and controversial aspects of HAB research relates to
312 methods to directly control or suppress blooms (Anderson 1997). Of the many
313 methods that have been proposed, removal of HAB cells through clay flocculation is

314 seen by some as promising in terms of efficiency, cost, and environmental impacts
315 (e.g., Sengco and Anderson, 2004; Lee et al. 2008). There are, however, those who
316 feel that the environmental impacts of this approach are unacceptable, or poorly
317 understood. In addition to the possible adverse ecological impact caused by the
318 addition of large amount of exotic materials (Shumway et al, 2003), other concerns
319 expressed relates to the constituents in the clay, which might include nutrients such as
320 phosphorus, or toxic or harmful metals and radioactive materials bound to the clay. As
321 an alternative to clays, sands are relatively inert or refractory and thus may minimize
322 these impacts. Most importantly, as a native part of the ecosystem, beach sand is
323 ecologically safe to the marine system which may avoid the fundamental concern
324 associated with clays. Large-scale dredging and beach nourishment projects abound in
325 nearshore waters worldwide, suggesting that environmental opposition to HAB
326 mitigation efforts using local sands might be minimal. In cases where beach sands
327 need to be conserved, commercially available sands may also be safe, cheap and
328 easily available to be used.

329 The modification technique using chitosan and PAC can not only turn local
330 beach sands or local soils into highly effective flocculants in the mitigation of HABs
331 in seawater, but is also useful in reducing the loading of sands/soils required for
332 effective cell removal, which is crucial for large scale field applications. Chitosan, a
333 commercially available product of edible food additives, is known to be a
334 biodegradable and non-toxic natural polymer. Compared with other chemical reagents,
335 chitosan is environmental friendly, but it might be a source of oxygen demand as it

336 decays. The amount of chitosan used is, however, much less than the amount of algal
337 biomass being sedimented, so this is not a serious concern. Nevertheless, it may be
338 worthwhile to develop techniques that could carry and release oxygen with the flocs
339 to combat this potential problem (Pan et al., 2009). In some coastal areas, it is also
340 possible to sink the algal blooms into the bottom and cover them using a second layer
341 of sands or local soils so that the cells can be permanently buried and sealed in the
342 sediment and turned into fertilizers for the growth of seaweeds, as Pan et al (2011)
343 demonstrated in shallow lakes. By decomposing the algal cells and the modifiers and
344 converting them into the biomass of seaweeds, the harmful blooms may be turned into
345 useful resources for the improvement of the ecosystem. However, this possibility
346 needs further study in marine systems affected by HABs. Although PAC (a compound
347 used in drinking water treatment) was needed to maintain the netting and bridging
348 function of chitosan in seawater, the adverse ecological effects of this compound in
349 seawater remain a concern. More research is needed in this area before larger-scale
350 applications can be undertaken. Similarly, efforts are needed to identify new,
351 environmentally benign modifiers that could replace PAC in this bloom control
352 strategy.

353

354 **5. Conclusion**

355 Dispersal of sands or local soils modified with chitosan and PAC achieved high
356 removal efficiency of marine HAB cells in a short time and prevented the escape of
357 significant numbers of motile organisms from the algal flocs. This method greatly

358 reduces potential environmental impacts by using relatively inert or refractory sand or
359 local and by using a biodegradable polymer such as chitosan, but there may be
360 environmental concerns about the use of PAC. With some additional studies, this
361 approach shows great promise to become an effective and environmentally acceptable
362 strategy for HAB mitigation.

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471 Figure Captions

472 Fig. 1. Algal removal efficiency of 100 mg L⁻¹ local sands, 100 mg L⁻¹ silica sands,
473 modifiers (10 mg L⁻¹ chitosan plus 10 mg L⁻¹ PAC), modified local sands (10
474 mg L⁻¹ chitosan plus 10 mg L⁻¹ PAC plus 100 mg L⁻¹ local sands) and
475 modified silica sands (10 mg L⁻¹ chitosan plus 10 mg L⁻¹ PAC plus 100 mg
476 L⁻¹ silica sands) at different time. (A) *A. carterae*, (B) *Chlorella sp.*

477 Fig. 2. Synergistic effect of chitosan and PAC on algae removal. (A) *A. carterae*,
478 (B) *Chlorella sp.*

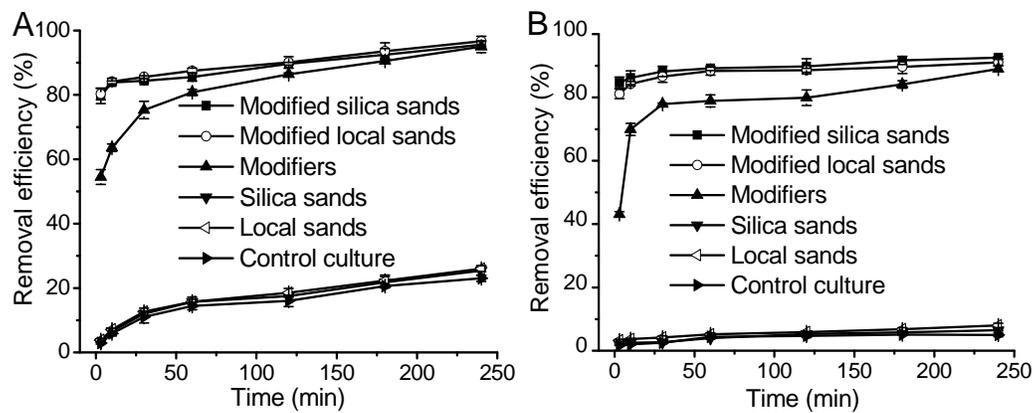
479 Fig. 3. Synergistic effect of chitosan and PAC on algal flocs. (A) Floc size, (B) Floc
480 size distributions at 7 min

481 Fig. 4. Synergistic effect of chitosan and PAC on algae viability. (A) with light and
482 added nutrients, (B) with no light or added nutrients

483 Fig. 5. Algal flocs micrographs with the magnification of 50 times. (A) Chitosan and
484 *A. carterae*, (B) PAC and *A. carterae*

485 Fig. 1.

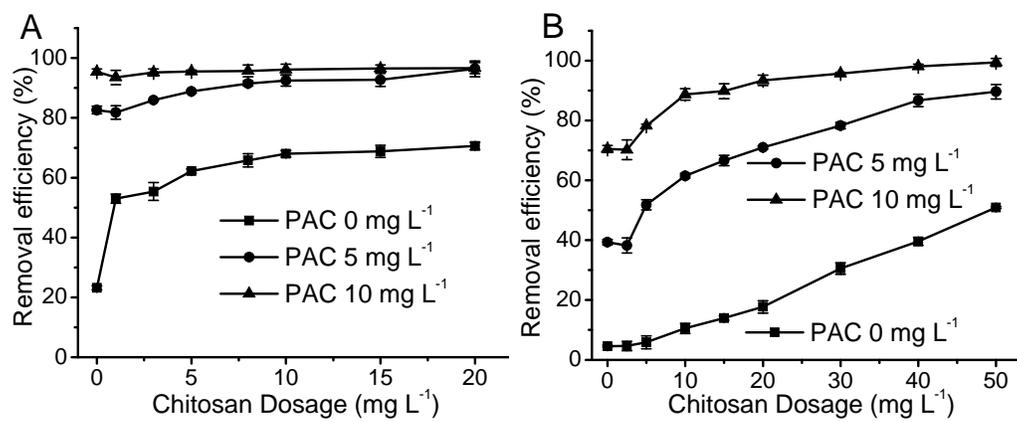
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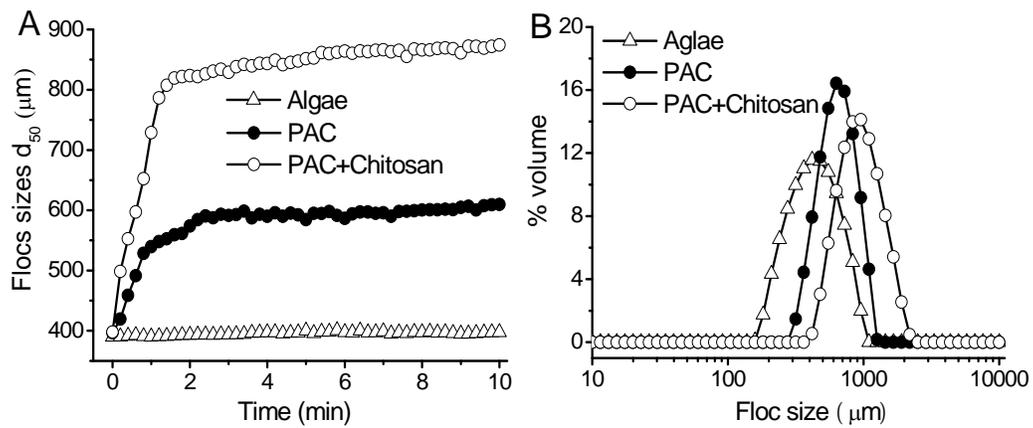
488 Fig. 2.

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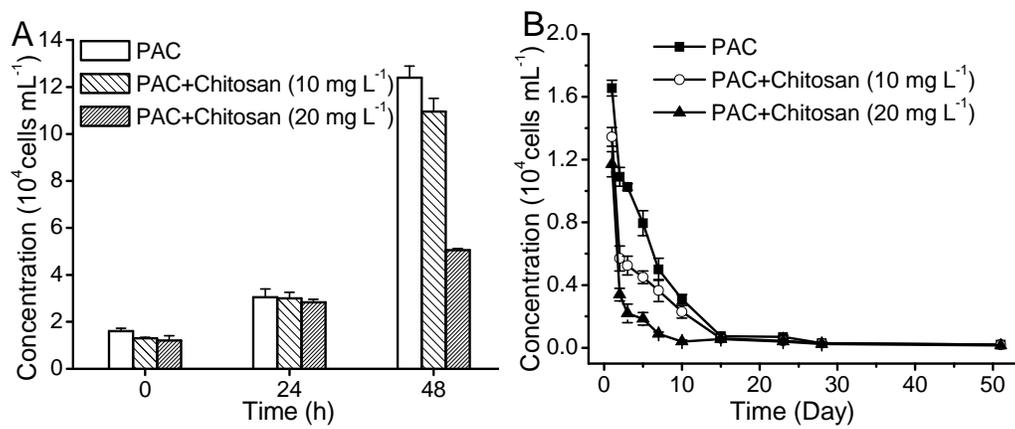
491 Fig. 3.



492

493 Fig. 4.

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495

496 Fig. 5.

